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Ecosystem cadastre of plant-soil interactions with nonferrous metals

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Abstract

Especially in high concentrations, heavy metals are potential toxic substances in the environment with untoward impacts on fauna and flora. The metal toxicity is enhanced by chemical associations. The immediate and remote concerns for sustainable growth and development are the identification, control and management of factors which affect plants and soils with concomitant hunger, quality food deprivation, impairment of health in humans and animals. It becomes imperative to create opportunities for sustainable mechanisms or modalities to eradicate the sources of the issues and challenges in heavy metal or trace element poisoning in the ecosystem cadastre.

Keywords: mining operations, environmental monitoring, environmental variability, spatiotemporal variations, environmental management

Introduction

The major challenges, issues and opportunities facing the agriculturist, environmentalist, or ecologist for sustainable management of the ecosystem and to combat hunger (Chukwuma 2020a) are the identification of factors which may contribute to growth inhibition or retardation of the fauna and impairment of human and animal health (Chukwuma, 1994a; Chukwuma 1995a; Chukwuma 2011; Sethy and Ghosh, 2013; Chukwuma 2016) by mining and mining-related operations as well as other anthropogenic activities and natural occurrences; and to find

remedies to resultant problems or issues (Chukwuma 2021). In polluting and derelict mine precincts including smelter sites, the waters become sterile due to the presence of toxic substances and infertile wastes. Certain of these toxic compounds are trace elements which are essential nutrients for higher plants, whereas at certain levels they may reach toxic concentrations. Disparate plant species accumulate several elements, particularly in mining areas via several mechanisms, such as competition and/or adaptive tolerance. Heavy metals constitute potential toxic substances and environmental problems; and present untoward

impacts on agriculture (Jamil et al., 2014), quality of life and the human environment; and the metal toxicity is inextricably-linked to the chemical associations in soil (Sivapatham et al., 2014). Superimposed on these are the environmental and public health implications of low level Pb exposure in the paediatric population which may culminate in psychometric deficits (Chukwuma, 1997; Chukwuma, 2020b). Cadmium is a persistent and ubiquitous environmental pollutant that is associated with renal dysfunction and different disorders (Adams et al., 2011). Trace element prevalence in plants growing on diverse unpolluted soils depicts widespread variation in defined elements (Kabata-Pendias, 2000); and this is observable in contaminated and polluted sites. Information on toxicity and heavy metal levels in crops is also pertinent to identify adequate, sub-adequate, and marginal intake contents for human and/or animals, so that diseases associated with trace element deficiency can be prevented, controlled or managed.

Issues, challenges and opportunities

The monitoring, evaluation and assessment of the trend and spatial dissemination of both natural and anthropogenic sources of trace elements and heavy metals in our environment are crucial to obtain and maintain database of the proportion of chemical elements and compounds in soils and plants. Baseline data for the occurrence and distribution of heavy metals as contaminants and pollutants are important as criteria for assessment of critical heavy metal contents in agricultural soils. Varied studies of the sources, fluxes and pathways of elements in the surface environment has been the focus of both national and international research communities in response to pollution and health impacts. Rural and agricultural areas have been precincts of interest where there may be incidence and prevalence of the perturbation of heavy metals on ecological and farming systems. Appreciable number of inhabitants is in the urban areas where natural geochemical elemental processes and in-puts undergo significant modifications and variations or completely masked by contamination due to

industrial activities and urban development. In several non-industrialised or low middle income countries with non-existent, non-compliant, unenforceable legislation or at its infancy, polluted or contaminated derelict mines or wastelands have been converted for agricultural purposes (Chukwuma, 1996).

Environmental toxicology is associated with heavy metals or trace elements, ecological and health impacts of environmental contaminants and pollutants as well as the factors related to weather and climatic conditions. Environmental modification, variation and geographical perspective are pertinent in toxic-element-related issues which interact on the food chain as evident in the Enyigba-Abakaliki lead and zinc mine precinct and several parts of the world (Chukwuma 1993a; Chukwuma, 1994a,b). This study tends to explicate or elucidate plant-soil interactions as inextricably-linked to the environment, agriculture and health.

Exploitation precinct are polluted by the mining operations and subjected the fauna and flora population to health and environmental risks via the food chain. Warm, dry climates and, to a certain magnitude, humid, hot climates induce upward translocation of trace elements in soil profiles (Kabata-Pendias and Pendias, 1992).

Despite the sampling period, metal levels in soils and plants are greater than background values (Bidar et al., 2009). In contradistinction to certain heavy metals which accumulate mostly in senescent leaves and leaf sheaths, Zn ostensibly has a uniform distribution in the entire plant.

The relative Pb uptake of rice seedlings was greatest in the pH range of 4.5-5.9, and least in the pH range 5-6.5 (Chukwuma, 1994c). Pb content of rice grain is, therefore, highly dependent on the soil pH and is the highest at mean pH 5.3-5.9. Rice grain uptake of Cd was highest in the pH range 5.4-5.9, and lowest in the pH range 4.5-6.5 (Chukwuma, 1994c). Cd content of rice grains is markedly dependent on the soil pH and is the maximum at pH 5.5 (Bingham et

al., 1980). Cd accumulates less in leaves than in lower parts; but was not evident in rice leaves to which ¹⁰⁹Cd was introduced (Kitagishi and Yamane, 1981). There is a perspicuous retardation of Pb translocation from root to shoot, as merely a small quantity is translocated to the shoot.

Heavy metal levels in soils and plants in the transect of a copper-tungsten mine demonstrated the impact of erstwhile base metal mining on the surficial ambient. Metal levels in soils depreciated with distance farther from the mine, and was manoeuvred mainly by aquatic mobility and topography (Jung, 2008). As determined (Chukwuma 1993b), metal contents in leaves were higher than in grains. The factors associated with the bioavailability of metals include soil pH, cation exchange capacity, organic matter content, soil texture and chemical interactions (Jung 2008).

Plants containing Zn at phytotoxic concentrations are not implicated in any toxicity in animals and humans because its adverse effect is attenuated in the food chain. Zn phytotoxicity may result at pH 6.5 or less. Due to the antagonistic relationship of Cd and Zn, the Cd/Zn ratio (0.002-0.004) in soils is ostensibly beneficial in the prevention or obliteration of toxic accumulation of Cd in food crops (Chukwuma 1993a; Chukwuma, 1994a,b).

Certain repercussions exist in the interaction of Cd with other metals, especially Zn. Cd and Zn are in the same group and share similar chemical attributes. They both exhibit biochemical antagonism; and their ratios are critical in biochemical outcomes. This similarity grants the latitude for their usual occurrence together with Zn predominating or in abundance. The established Cd:Zn ratio in most staple foods is circa 1:100 (National Academy of Science, 1977). Data from a derelict mine transect revealed that the Cd:Zn ratios were 1:100 and 1.57 respectively, for cultivated (staple foods) and wild plant species (Chukwuma, 1993a). It is suggested that there is mitigation in the potential toxicity of Cd by Zn via simple mass action effect

specifically for cultivars, while other superimposed tolerance or adaptive mechanisms are operative in the wild plants.

The presentation of toxic elements in the soil-plant system correlates to their discrete states. Elemental toxicity as determined by plant uptake is associated with the outcome both in the liquid and solid states of the soil. Toxic elements have stronger adsorption at high pH with more significant effect in light textured soils. Indubitably, pH variations influence trace metal solubility, as they are inter alia also governed by reductive dissolution and organic acid production (Charlatchka and Cambier, 2000). In a mine transect, the mean soil pH was 6.01, and light textured brown powdery soil had the highest Zn and Cd concentrations (Chukwuma, 1994b).

There was disparity of the energy binding of each element to the soil. It was easier for Cu and Zn to be extracted from soil solution than Cd and Pb (Huang et al., 1977). There were fluctuations of metal levels in plant parts in contrast to soils over time. Bioaccumulation and transfer factors ranges farther away reflect and depict higher soil load of chemical elements at the mine and its periphery. The findings were that *Panicum maximum* is the best indicator species for Cd and Pb for the investigated soils (Chukwuma, 1994c). The most appropriate indicator plant species for trace element levels for the investigated cultivated areas were grass (Cu), yam leaves (Cu, Ni), rice leaves (Mn), rice leaves and grains (Ni) and cassava leaves (Zn). The greatest range was described for Mn in yam leaves, and the least variation for Ni in rice leaves (Chukwuma, 1995b).

Imperata cylindrica accumulated Pb but excluded both Zn and Cd. The *Ageratum conyzoides* was determined as either an accumulator or indicator depending on the chemical element or whether it was being classified as regards plant uptake or the plant/soil ratio system (Chukwuma, 1993b). *Thlaspi calaminare*, a variant of the mountain pansy, *Viola lutea* accumulated >10%, whereas *Equisetum arvense* took up 0.1 to 1% Zn, and other trace elements in Western Europe

(Edington and Edington, 1977). The basil *Becium homblei* is associated with copper deposits in Africa (Edington and Edington, 1977). The grass, *Eriachne mucronata* is associated with lead uptake (Cannon, 1971) in Australia. In Poland, dandelion, orchard grass, plantain, and lichens are ostensibly the best indicator plants for Cd levels in soils (Kabata-Pendias and Dudka, 1990). To elucidate the food-chain linkage in the bioaccumulation of heavy metals, the goat weed, *Ageratum conyzoides* may serve as a better indicator of the distribution of Zn, Pb and Cd. *Imperata cylindrica* and *Bryophyllum pinnatum* may be tenable for biopurification purposes. *Dioscorea bulbifera* could be useful for biological locating of ores (Chukwuma, 1993b). The mungbean may impose as a nickel indicator, accumulator, or excluder (Ahmad et al., 2007).

The usage of plants for monitoring heavy metal pollution in the terrestrial environment must inculcate the complex, integrated impacts of pollutant and contaminant sources as well as soil-plant variables. In order to be detectable in plants, it becomes critical that pollutant and contaminant sources markedly spike the plant available metal level in soils. The essential and non-essential heavy metals Fe, Ni, Mn, Zn, Cu, Cd, Cr) and Pb analyzed in four selected medicinal plants, *Capparis spinosa*, *Peganum harmala*, *Rhazya stricta*, and *Tamarix articulata* (Shah et al., 2013) showed that these plants are widely consumed as traditional medicine for treatment of diverse ailments. The heavy metal concentrations in the plants was found to decrease in the order: Fe > Zn > Mn > Cu > Ni > Cr > Cd > Pb. These medicinal plants accumulate these elements at varied levels. Monitoring such medicinal plants for heavy metal contents is immensely valuable to physicians, health planners, health care professionals, and policymakers for the protection of the public from their adverse toxicity impacts.

Microorganisms which are capable of resisting elevated concentrations of toxic heavy metals are effective agents of bioremediation regarding such contaminants and pollutants. Halophilic microorganisms are useful in bioremediation due

to their efficient and effective removal of heavy metals (Sowmya et al., 2014) in soil contaminated with various levels of Ni. The rates of germination and percentage significantly diminished in the plants (Jamil et al., 2014). The pigments of photosynthesis, chlorophyll a, chlorophyll b, and carotenoids also when different Ni concentrations were applied to the soil. Total protein and organic nitrogen decreased at elevated Ni levels. Bacteria serve in competent phytoremediation of heavy metal deranged soil. Improvement in seed germination and biochemical attribute of plants following Ni stress were detected with the inoculation of *Bacillus licheniformis* NCCP-59 strain. This bacterial strain may protect plants from the debilitating impacts of Ni, and useful for the phytoremediation of Ni contaminated soils.

The reclamation of nonferrous metal-polluted soil via phytoremediation requires an entirely permanent plant cover. Selection of the most appropriate plant species necessitates the investigation of chemical element effects on plants over the specified period, and assurance that the chemical elements remain stored in root systems and not transferred to the aerial aspects. It then becomes imperative to study the seasonal and annual variations of metal bioaccumulation, transfer, and phytotoxicity (Chukwuma, 1993a; Bidar et al., 2009). Investigations of the spatiotemporal variations of sediment quality containing selected heavy metals and total hydrocarbons (THC) with the application of multivariate statistical techniques, such as principal components, cluster and lineal discriminant analyses discerned the environmental interpretation of a limited dataset for the identification of environmental quality (Alvarez et al., 2014). It was determined that Zn, Cu, Pb, V, As and THC significantly contributed to sediment quality variations during the sampling period.

Environmental variability and resource dynamics

Soil constitutes an aspect of a vital environmental, ecological and agricultural resource that needs protection from further degradation. An optimum supply of healthy food (Chukwuma 2020a; Chukwuma 2021) is required for the ever-increasing global population. Chemical elements can affect both the yield and composition of crops. Therefore, a determination of the elemental status of all cultivated lands has to be made for the identification of yield-limiting deficiencies of essential micronutrients and the corresponding contaminated and polluted soils. The extensive variability of soil composition is influenced by several factors, the most vital of which are climatic status and parent material. Research on the association between trace elements and soil constituents explicates their behaviour in soils and in plant uptake. The fate and pattern of mobilized chemical elements by dissolution of parent substances are dependent on the ionic speciation of the soil solution. Significant alterations in element concentrations in soil solutions result due to precipitation, evaporation, and plant transpiration. The trajectories of trace element distribution, translocation and accumulation vary expansively for any element, plant species, and growth season.

Environmental pollution data depicting environmental variability vary extensively and are susceptible to disparate uncertainties, such as inter alia distance, pollution sources, natural background variation, spatiotemporal pollution build-up or degradation. Environmental variability relates to the unique variation in pollution levels between population units. In essence, the levels, dynamics and significance of heavy metals in composted waste materials are critical from two potentially conflicting dimensions of environmental legislation as to: (a) explicate and define end-of-waste criteria and augment recycling of composted residuals on land and (b) protect and preserve soil quality as well as create a barrier to contamination and pollution (Chukwuma 1996; Smith, 2009).

The development of a monitoring programme for the sampling of the same plant species using standardized sampling and analytical methods for the maximization of the comparability of data is pertinent (Chukwuma, 1998). There are expansive highly metalliferous precincts with ponds, lakes, rivers, streams, and other aquatic resources (Chukwuma 2020c). with diverse fauna and flora which are vulnerable to extreme quantities of trace elements or heavy metals which may be inimical to the ambient, animal and human health. Most of these plants are ubiquitous, and could be sampled regularly with standardized sampling and analytical techniques to monitor the time-trend or spatiotemporal dissemination of chemical elements which are potential environmental, animal and human health hazards.

Discussion

Invariably, there is the overwhelming need for the simultaneous management of soil on the entire spectrum of essential and toxic elements. Crops and livestock usually become subjected to latent hunger or latent toxicities. Plant response to an element or nutrient may be suppressed when growth or development is inhibited by toxicity or deficiency of another element antagonistically or synergistically. Soils with low levels of contaminants need protection using stringent measures from superimposed contamination, that is, pollution. Soils polluted by heavy metals must either be cleaned or at the least, elemental removal conducted to meet minimal concentration requirements for human health, optimum growth for plant and soil micro- and macro-organisms. Observed elemental concentrations in a given soil which exceed normal values, may be indications of untoward in-puts by agricultural activities, emissions via the air, parent rocks, and transportation of accumulated pollutants to soils, waters and plants (Gupta, 1991; Chukwuma, 2014).

Although, it is not every ambient that has been adversely exposed and susceptible to heavy metal and trace element contamination, pollution and toxicity, there is pertinence for concern (Guan and Peart, 2006; Chukwuma, 2014). Pb levels were

found to be substantially higher in tree leaves on roadsides than in a park. Heavy metal concentrations were lower in the roots than in leaves; thus suggesting that heavy metal pollution of trees was mainly due to air pollution (Guan and Peart, 2006).

it is pertinent to address environmental perturbations in all aspects. Mobilization of different chemical elements at toxic levels in the environment can accumulate in the food chain. There is need for continuous investigation on the effects of mobility of metals for the evaluation of pollutants and contaminants to be absorbed by the biota and transferred to groundwater. This will provide for a replicable method for the assessment of aquatic ecosystem health in adjacent plateau reservoirs (Wu et al., 2014) and mine precinct. also, bioaccessibility tests are valuable on contaminated and polluted soils to evaluate phytostabilization impacts on the exposition risks for fauna and flora (Chukwuma, 2014).

Conclusion

Trace elements and their compounds constitute vital components of the biotic environment, and a vast majority of them are essential for the health of the fauna and flora. However, essential chemical elements can accumulate to toxic concentrations in the environment via anthropogenic activities, such as non-ferrous metal mining operations. It is pertinent to monitor the concentrations of elements which are bioaccumulated in wild terrestrial plant species for monitoring purposes. Certain plants specifically take up select trace elements in metal-contaminated or polluted soils, and have been used to locate ores. Data acquisition and collation using well-defined monitoring programme regarding the spatiotemporal distribution of metal- contaminants in a particular plant species, particularly in ambients with similar plant and soil features.

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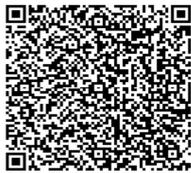
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